



REVIEW

Seaweed as a feed additive to mitigate enteric methane emissions in ruminants: Opportunities and challenges

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Highlights

- *Asparagopsis taxiformis* has demonstrated significant potential in reducing enteric methane emissions in ruminants, achieving reductions of up to 99% *in vitro* and *in vivo*, primarily due to its bromoform content.
- The application of seaweed as a feed additive faces challenges, including potential heavy metal contamination, environmental risks associated with bromoform, and the need for a sustainable cultivation and processing supply chain.
- Further research is needed to identify low-bromoform seaweed species and investigate additional bioactive compounds to optimize methane mitigation strategies.

Abstract

Cutting farming-related methane emissions from ruminants is critical in the battle against climate change. Since scientists initially investigated the potential of marine macroalgae to reduce methane emissions, using seaweeds as an anti-methanogenic feed additive has become prevailing in recent years. *Asparagopsis taxiformis* is the preferred species because it contains a relatively higher concentration of bromoform. As a type of halogenated methane analogue, bromoform contained in *A. taxiformis* can specifically inhibit the activity of coenzyme M methyltransferase, thereby blocking the ruminal methanogenesis. However, bromoform is a potential toxin and ozone-depleting substance. In response, current research focuses on the effects of bromoform-enriched seaweed supplementation on ruminant productivity and safety, as well as the impact of large-scale cultivation of seaweeds on the atmospheric environment. The current research on seaweed still needs to be improved, especially in developing more species with low bromoform content, such as *Bonnemaisonia hamifera*, *Dictyota bartayresii*, and *Cystoseira trinodis*. Otherwise, seaweed is rich in bioactive substances and exhibits antibacterial, anti-inflammatory, and other physiological properties, but research on the role of these bioactive compounds in methane emissions is lacking. It is worthy of deeper investigation to identify more potential bioactive compounds. As a new focus of attention, seaweed has attracted the interest of many scientists. Nevertheless, seaweed still faces some challenges as a feed additive to ruminants, such as the residues of heavy metals (iodine and bromine) and bromoform in milk or meat, as well as

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the establishment of a supply chain for seaweed cultivation, preservation, and processing. We have concluded that the methane-reducing efficacy of seaweed is indisputable. However, its application as a commercial feed additive is still influenced by factors such as safety, costs, policy incentives, and regulations.

Keywords: seaweed, *Asparagopsis taxiformis*, bromoform, methane emission, ruminant

1. Introduction

Methane has a global warming potential (GWP) of 86 or 28 times greater than CO₂ on a 20- or 100-year time scale, respectively (IPCC 2021). Agriculture contributes to nearly half of the global anthropogenic methane production, followed by fossil fuels and waste (Opio *et al.* 2013). The livestock sector is among the most significant contributors to methane emissions, especially ruminant production. Numerous approaches have been proposed, with dietary manipulation being one of the most effective ways to mitigate enteric methane emissions from ruminants (Arndt *et al.* 2022). Seaweeds (macroalgae) are marine plant resources extensively used in human foods, natural cosmetics, pharmaceuticals, and nutraceuticals (Holdt and Kraan 2011; Rengasamy *et al.* 2020). Interest in the application of macroalgae to mitigate enteric methane emissions from ruminants has increased dramatically in recent years (Machado *et al.* 2016a; Antaya *et al.* 2019; Roque *et al.* 2019; Kinley *et al.* 2020; Vijn *et al.* 2020). Although bromoform is believed to be the primary active ingredient responsible for methane mitigation (Machado *et al.* 2016b), other bioactive compounds such as phlorotannins, alkaloids, and peptides may also play a role in reducing enteric methane emissions (Abbott *et al.* 2020). This review summarizes recently published studies using seaweed as an anti-methanogenic feed additive, discusses the challenges and limitations, and proposes perspectives for future research.

2. Rumen methanogens and mitigation of methanogenesis

2.1. Pathway of methanogenesis in the rumen

Ruminants rely on rumen microbes to ferment dietary components into volatile fatty acids (VFAs), the major energy source for the host (Moss *et al.* 2000). During fermentation, the hydrogenases in bacteria or protozoa transfer electrons from reduced cofactors to H⁺ forming H₂, which is further transferred to methanogens and used for methanogenesis (Pitta D *et al.* 2022). Primary

substrates can be used by rumen methanogens to produce methane including CO₂, methyls (methanol, methylamines), and acetate (Fig. 1). Based on substrate utilization, rumen methanogenesis is divided into 3 pathways, including hydrogenotrophic, acetoclastic, and methylotrophic methanogenesis (Berghuis *et al.* 2019). In hydrogenotrophic methanogenesis, rumen methanogens reduce CO₂ to methane using H₂ as an electron donor through the Wolf cycle (Thauer 2012). Briefly, CO₂ binds initially to methanofuran (MFR) and is subsequently reduced to formylmethanofuran (CHO-MFR) by formylmethanofuran dehydrogenase (Fmd) and Fd_{red} (Wagner *et al.* 2016). The formyl group is then transferred to tetrahydromethanopterin (H₄MPT) to form CHO-H₄MPT, which is subsequently dehydrated to CH₃-H₄MPT with reduced F₄₂₀ (F₄₂₀H₂) as the electron donor (Thauer 1998). The methyl group of CH₃-H₄MPT is transferred into HS-CoM to form CH₃-S-CoM, a typical central intermediate shared by all 3 methanogenic pathways. Finally, CH₃-S-CoM is converted to methane using HS-CoB as the electron donor (Wongnate *et al.* 2016). In acetoclastic methanogenesis, methanogens convert acetate to acetyl-coenzyme A (CH₃-CO-S-CoA) using adenosine triphosphate (ATP) and HS-CoA. Acetyl-coenzyme A is split to form enzyme-bound methyl and carbonyl groups (CO-S-CoA). The methyl group is then transferred to H₄MPT, which follows the last 2 steps of the hydrogenotrophic pathway. Following the methyl-coenzyme M reductase-catalyzed reaction, the carbonyl group is oxidized to CO₂ with Fd_{ox} as the electron acceptor. Methylotrophic methanogens utilize methylated compounds including trimethylamine, dimethyl sulfate, and methanol (Enzmann *et al.* 2018). The methyl group from methyl-containing compounds (CH₃-R) is transferred to a cognate corrinoid protein and then to CoM-SH in the methylotrophic pathway (Lyu *et al.* 2018).

2.2. Compounds that potentially mitigate rumen methanogenesis

The conversion of methyl-CoM to methane is the final step in all 3 methanogenic pathways. This step is catalyzed by methyl-coenzyme M reductase (MCR), an enzyme found in all methanogens in which nickel is bound to cofactor F₄₃₀.

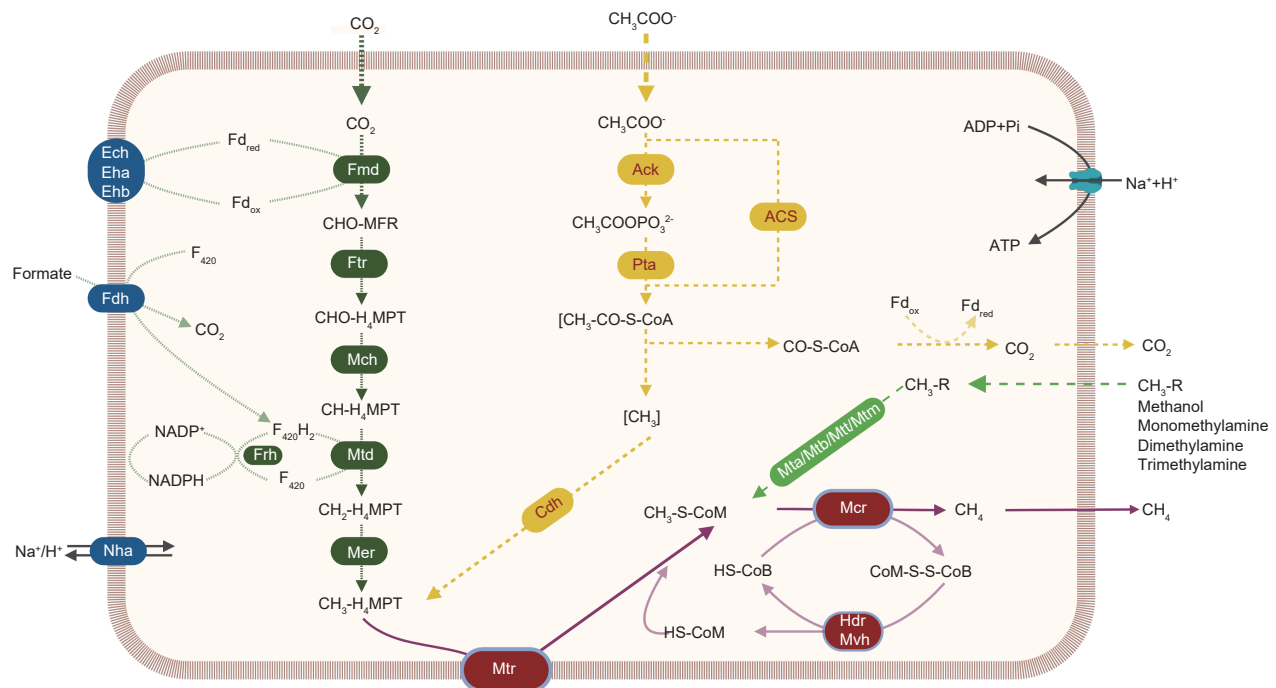


Fig. 1 Three pathways of methanogenesis. Hydrogenotrophic methanogenesis is the most widespread pathway which uses CO_2 and H_2 to produce methane. Acetoclastic methanogenesis splits acetate into methane and CO_2 . Methylotrophic methanogenesis is a group of methanogens that can use reduced one-carbon compounds, such as methanol, monomethylamine, dimethylamine, and trimethylamine. Frh, F_{420} -reducing hydrogenase; Nha, Na^+/H^+ antiporter; Fmd, formylmethanofuran dehydrogenase; Ftr, formylmethanofuran-tetrahydromethanopterin N-formyltransferase; Mch, methenyltetrahydromethanopterin cyclohydrolase; Mtd, 5,10-methylenetetrahydromethanopterin dehydrogenase; Mer, 5,10-methylenetetrahydromethanopterin reductase; Ack, acetate kinase; Pta, phosphotransacetylase; ACS, acetyl coenzyme A synthetase; Cdh, CO dehydrogenase/acetyl-CoA synthase; Mta, methyltransferase; Mtb, methanol cobalamin methyltransferase; Mtt, trimethylamine methyltransferase; Mtm, monomethylamine methyltransferase; Ech/Eha/Ehb, energy-converting hydrogenase; Fdh, formate dehydrogenase; Mtr, coenzyme M methyltransferase; Mcr, methyl coenzyme M reductase; Hdr, heterodisulfide reductase; Mvh, F_{420} -non-reducing hydrogenase.

a tetrapyrrole derivative (Ermler *et al.* 1997; Wongnate *et al.* 2016). The structural analog of methyl-coenzyme M that binds to the active site of methyl-coenzyme M is one of the most promising enteric methane inhibitors (Fig. 2). Various halogenated, sulfonated compounds, such as 2-bromoethanesulfonate (BES) (Robert *et al.* 1978; Gräwert *et al.* 2014), 2-chloroethanesulfonate (CES) (Ungerfeld *et al.* 2004), and 3-bromopropanesulfonate (BPS) (Ungerfeld *et al.* 2004), are structural analogues of CoM, and they can competitively and specifically inhibit MCR activity, decreasing methane production at relatively low concentrations (Ellermann *et al.* 1989; Gräwert *et al.* 2014; Patra *et al.* 2017). In addition, 3-nitrooxypropanol (3-NOP) has a similar molecular structure to methyl-coenzyme M, which targets the nickel enzyme of MCR (Duin *et al.* 2016; Pitta D W *et al.* 2022). It has been reported that 3-NOP reduces methane emissions in ruminants by over 30% on average (Melgar *et al.* 2021).

Coenzyme M methyltransferase (MTR) is another essential enzyme in methanogenesis and halogenated methane analogues (HMAs) could potentially inhibit its

activity by interfering with the cobamide-dependent methyltransferase step of methanogenesis (Fig. 2) (Bauchop 1967; Wood *et al.* 1968; Tomkins *et al.* 2009). The HMAs including bromoform and dibromochloromethane could inhibit methane production even at low concentrations (Machado *et al.* 2016b). Bromoform, a naturally occurring halogenated volatile organic compound derived from chemical and biological sources such as marine phytoplankton and macroalgae, is believed to be the primary active ingredient in seaweeds responsible for methane mitigation (Machado *et al.* 2016b).

Besides bromoform, seaweeds contain abundant bioactive compounds, such as carbohydrates and phlorotannins, alkaloids and peptides. These bioactive compounds may also play a role in reducing enteric methane (Abbott *et al.* 2020). For example, phlorotannins can act as H_2 sinks and reduce the H_2 concentration in the rumen (Kim *et al.* 2022). It is hypothesized that the presence of these metabolites in the rumen reduces the availability of H_2 , thereby reducing the substrate available for methanogenesis.

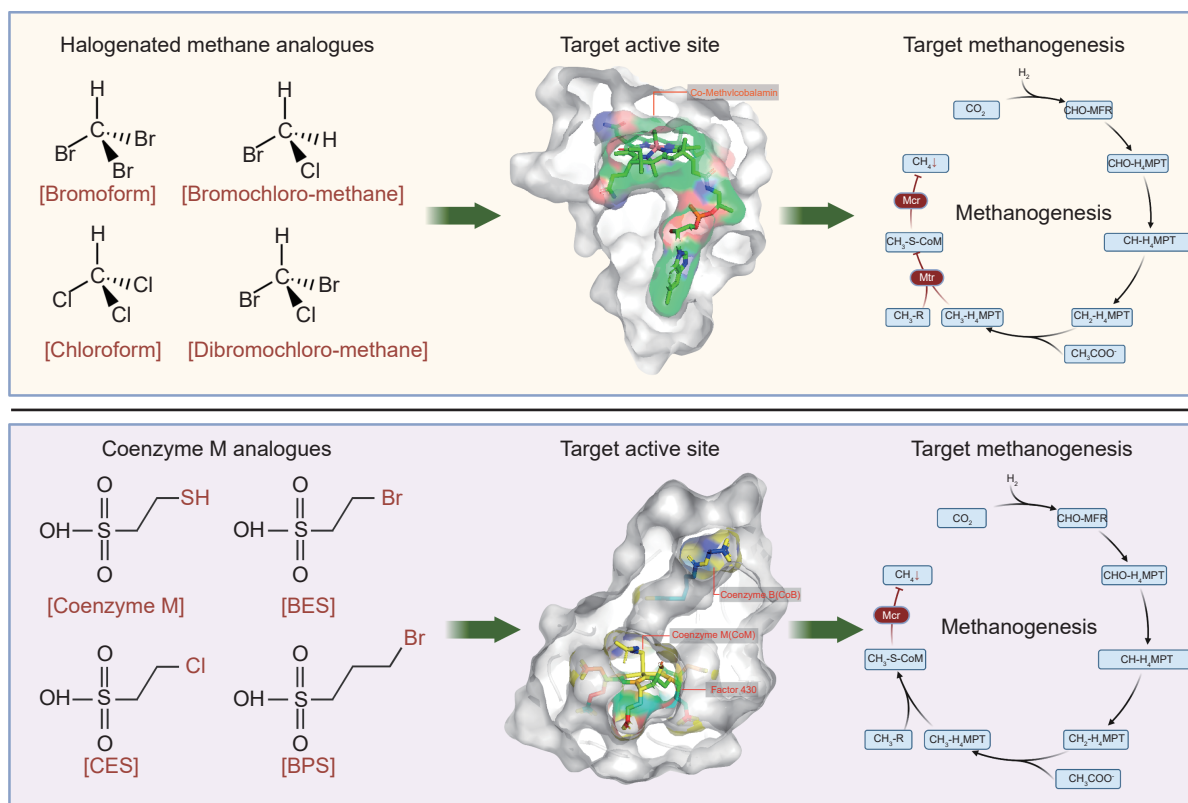


Fig. 2 The chemicals of the halogenated methane analogue and coenzyme M analogue target key enzymes in the methanogenesis pathway. BES, 2-bromoethanesulfonate; CES, 2-chloroethanesulfonate; BPS, 3-bromopropanesulfonate.

3. Seaweeds affect the methane mitigation of ruminant animals

There has been growing interest in feeding seaweeds to ruminants in recent years because of their significant effects on methane mitigation. The seaweeds used in previous studies mainly belong to red (Rhodophyta), brown (Phaeophyceae), and green algae (Chlorophyta).

3.1. Methane mitigation potential of red algae

Red algae, with more than 6,100 species, are essential coastal ecosystem components and economically significant as natural foods and a source of gelling agents (Guiry 2012). Their color results from the presence of pigments, phycoerythrin, and phycocyanin. Red algae are the most studied seaweeds in ruminants, particularly *Asparagopsis taxiformis*, and exhibit strong methane mitigation effect both *in vitro* (66–99%) and *in vivo* (30–99%) (Table 1 and Fig. 3-A). Numerous studies demonstrated that even low levels (2–10% dietary organic matter, OM) of *A. taxiformis* supplementation could exhibit strong methane mitigation capabilities (Table 1). Mihaila et al. (2022) compared the effect of multiple red algae

on methane mitigation *in vitro*, and found that adding *Asparagopsis armata* at 2% OM completely inhibited methane production and significantly increased hydrogen production. In addition, *Bonnemaisonia hamifera* demonstrated the most significant reduction in methane (98.9%) and an increase in the production of hydrogen with the addition of 10% OM, while *Euptilota formisissima*, *Gracilaria vermiculophylla*, *Laurencia filiformis*, *Hypnea pannosa*, and *Plocamium cirrhosum* reduced methane production by 40 to 50% (Mihaila et al. 2022). Several *in vivo* studies also showed that *A. taxiformis* significantly mitigates methane emissions from ruminants (Table 1). For example, Li et al. (2018) observed an 80% reduction in methane production when *A. taxiformis* was added to the diets of sheep. Dietary supplementation with *A. taxiformis* (0.2 to 0.5 OM basis) decreased enteric methane production in beef steers and heifers by 67 to 98% (Kinley and Fredeen 2015; Roque et al. 2021; Almeida et al. 2022). Methane emissions from dairy cows decreased by 30–35% when *A. taxiformis* was supplemented (Stefenoni et al. 2021; Angellotti et al. 2022; Wasson et al. 2022a). In general, the results of *in vivo* experiments were consistent with those obtained *in vitro*, although factors such as dosage, experimental

Table 1 Summary of studies using red algae (Rhodophyta) for methane reduction, including methods, inclusion level, results, and references¹⁾

Seaweed	<i>In vivo</i> / <i>In vitro</i>	Animal	Dose	CH ₄ (Δ%)	TVFA (Δ%)	C2 (Δ%)	C3 (Δ%)	C2/C3 (Δ%)	DMI (Δ%)	References
<i>Asparagopsis taxiform</i>	<i>In vitro</i>	—	2% OM	99 ↓	11.5 ↓	37 ↓	68 ↑	63 ↓	—	Machado et al. (2016a)
<i>A. taxiform</i>	<i>In vitro</i>	—	5% DM	74 ↓	—	—	—	—	—	Brooke et al. (2020)
<i>A. taxiform</i>	<i>In vitro</i>	—	2% DM	66 ↓	7 ↓	8 ↓	17 ↓	10 ↑	—	Terry et al. (2022)
<i>A. taxiform</i>	<i>In vitro</i>	—	9 mg	99 ↓	—	—	—	—	—	Alvarez et al. (2022)
<i>A. taxiform</i>	<i>In vitro</i>	—	2% DM	81 ↓	—	—	—	—	—	Gunter et al. (2022)
<i>A. armata</i>	<i>In vitro</i>	—	2% OM	100 ↓	22 ↓	23 ↓	47 ↑	—	—	Mihaila et al. (2022)
<i>Bonnemaisonia hamifera</i>	<i>In vitro</i>	—	10% OM	98.9 ↓	25 ↓	21 ↓	53 ↑	—	—	Mihaila et al. (2022)
<i>Devaleraea mollis</i>	<i>In vitro</i>	—	2% DM	23 ↓	—	—	—	—	—	Gunter et al. (2022)
<i>Euptilota formisissima</i>	<i>In vitro</i>	—	10% OM	51 ↓	13 ↓	10 ↓	27 ↑	—	—	Mihaila et al. (2022)
<i>Gracilaria vermiculophylla</i>	<i>In vitro</i>	—	25% DM	38 ↓	5 ↓	3 ↑	5 ↑	11 ↑	—	Maia et al. (2016)
<i>Laurencia filiformis</i>	<i>In vitro</i>	—	0.2 g OM	40 ↓	12 ↓	3 ↑	0.6 ↓	1.6 ↑	—	Machado et al. (2014)
<i>Hypnea pannosa</i>	<i>In vitro</i>	—	0.2 g OM	43 ↓	ns	4 ↑	6 ↓	9 ↑	—	Machado et al. (2014)
<i>Plocamium cirrhosum</i>	<i>In vitro</i>	—	0.2 g OM	40 ↓	12 ↓	8 ↓	21 ↑	—	—	Mihaila et al. (2022)
<i>A. taxiform</i>	<i>In vivo</i>	Sheep	3% OM	80 ↓	29 ↓	16 ↓	54 ↑	45 ↓	—	Li et al. (2018)
<i>A. taxiform</i>	<i>In vivo</i>	Beef steer	0.5% OM	80 ↓	—	—	—	—	14.2 ↓	Roque et al. (2021)
<i>A. taxiform</i>	<i>In vivo</i>	Beef steer	0.2% OM	98 ↓	—	—	—	—	10.8 ↓	Kinley et al. (2020)
<i>A. taxiform</i>	<i>In vivo</i>	Beef cattle	0.4% OM	67 ↓	—	—	—	—	ns	Almeida et al. (2022)
<i>A. taxiform</i>	<i>In vivo</i>	Dairy cow	0.5% DM	34 ↓	11 ↓	6.6 ↓	8.5 ↑	12 ↓	7 ↓	Stefenoni et al. (2021)
<i>A. taxiform</i>	<i>In vivo</i>	Dairy cow	0.3% OM	35 ↓	—	—	—	—	7.6 ↓	Angellotti et al. (2022)
<i>A. taxiform</i>	<i>In vivo</i>	Red cow	0.5% OM	60 ↓	—	11 ↓	23 ↑	—	13.8 ↓	Krizsan et al. (2023)
<i>A. taxiform</i>	<i>In vivo</i>	Goats	0.5% DM	30–40 ↓	—	—	—	—	—	Romero et al. (2022)
<i>A. taxiform</i>	<i>In vivo</i>	Dairy cows	0.75% OM	30 ↓	—	—	—	—	22 ↓	Wasson et al. (2022a)
<i>A. taxiform</i>	<i>In vivo</i>	Dairy cows	1% OM	67.2 ↓	—	—	—	—	38 ↓	Roque et al. (2019)
<i>Chondrus crispus</i>	<i>In vivo</i>	Dairy cows	126–136 g d ⁻¹	ns	—	—	—	—	ns	Muizelaar et al. (2023)
<i>C. crispus</i>	<i>In vivo</i>	Dairy cows	6% DM	13.9 ↓	—	—	—	—	ns	Reyes et al. (2023)

¹⁾ CH₄, methane; TVFA, total volatile fatty acids; C2, acetate; C3, propionate; C2/C3, the ratio of acetate to propionate; DMI, dry matter intake; OM, organic matter; DM, dry matter. Δ%, treatment response over control; —, not quantified; ↑ and ↓ represent the percentage increase and decrease, respectively, of the value in the treatment group supplemented with seaweed compared with the control group. ns, not statistically significant.

period, and animal species may affect its efficacy.

3.2. Methane mitigation potential of brown and green algae

Brown algae have high concentrations of carotenoids, such as yellow-brown fucoxanthin. There are approximately 1,800 species of brown algae worldwide, and most are known to exist in marine water (Makkar et al. 2016). Most of the studies using brown algae were conducted *in vitro* (Table 2 and Fig. 3). Overall, the inhibitory effect of most of the brown algae on methane emission is weaker than that of red algae, which decreased methane production by less than 50%, except for *Cystoseira trinodis* (Dubois et al. 2013) and *Dictyota bartayresii* (Machado et al. 2014) that decreased methane production by 80 and 92%, respectively.

As for green algae, there are 2,200 currently known species, and their color results from the presence of chlorophyll (Makkar et al. 2016). Similar to brown algae, the methane inhibitory effect of green algae was basically

evaluated *in vitro* (Table 3). Specifically, *Cladophora patentiramea* was the most effective species that could reduce methane production by 66%. *Caulera taxifolia*, *Cladophora coelothrix*, *Chaetomorpha linum*, *Ulva* sp. and *Ulva ohnoi* could reduce methane production by 27 to 50% after 72 h of incubation (Machado et al. 2014).

The DNA-barcoding using the ribulose-1,5-bisphosphate carboxylase/oxygenase large subunit (*rbcL*) gene has proven its usefulness in studying seaweed phylogenetic diversity (Alshehri et al. 2019). The NCBI nucleotide submitted sequences of the *rbcL* gene of 23 seaweed species have been used to generate a phylogenetic tree (Fig. 3-B), which contains 3 clades, Chlorophyta, Ochrophyta, and Rhodophyta. We found that the close evolutionary relationships of seaweeds exhibited similar methane mitigation potential in only a few clades but not in other clades. These results indicated that seaweeds of similar species do not all exhibit the same methane inhibition potential, which may be related to the types and contents of methane-inhibiting compounds contained in the seaweeds. The bromoform content of seaweed was

Table 2 Summary of studies using brown algae (Ochrophyta) for methane reduction, including methods, inclusion level, results, and references¹⁾

Seaweed	<i>In vivo</i> / <i>In vitro</i>	Animal	Dose	CH ₄ (Δ%)	TVFA (Δ%)	C2 (Δ%)	C3 (Δ%)	C2/C3 (Δ%)	References
<i>Ascophyllum nodosum</i>	<i>In vitro</i>	–	5% DM	8.9 ↓	16 ↓	17 ↓	6 ↓	14 ↓	Künzel et al. (2022)
<i>Fucus vesiculosus</i>	<i>In vitro</i>	–	5% DM	3.6 ↓	19 ↓	22 ↓	3 ↓	21 ↓	Künzel et al. (2022)
<i>Colpomenia sinuosa</i>	<i>In vitro</i>	–	0.2 g OM	49 ↓	17 ↓	2 ↓	14 ↑	15 ↓	Machado et al. (2014)
<i>Cystoseira trinodis</i>	<i>In vitro</i>	–	160 mg OM	80 ↓	–	–	–	–	Dubois et al. (2013)
<i>C. trinodis</i>	<i>In vitro</i>	–	0.2 g OM	45 ↓	29 ↓	6.7 ↓	25 ↑	21 ↓	Machado et al. (2014)
<i>Dictyota bartayresii</i>	<i>In vitro</i>	–	0.2 g OM	92 ↓	39 ↓	4.8 ↓	41 ↑	32 ↓	Machado et al. (2014)
<i>Ecklonia radiata</i>	<i>In vitro</i>	–	10% OM	21 ↓	7 ↓	2.4 ↓	15.7 ↑	–	Mihaila et al. (2022)
<i>Hormophysa triquetra</i>	<i>In vitro</i>	–	0.2 g OM	44 ↓	24 ↓	1.5 ↑	10 ↑	7 ↓	Machado et al. (2014)
<i>Laminaria ochroleuca</i>	<i>In vitro</i>	–	25% DM	14 ↑	10 ↓	2.4 ↑	5.8 ↓	8.4 ↑	Maia et al. (2016)
<i>Saccharina latissima</i>	<i>In vitro</i>	–	–	80 ↓	–	–	–	–	Olijhoek et al. (2022)
<i>Sargassum flavicans</i>	<i>In vitro</i>	–	0.2 g OM	34 ↓	ns	3.8 ↑	4.4 ↓	8.6 ↑	Machado et al. (2014)
<i>Padina australis</i>	<i>In vitro</i>	–	0.2 g OM	50 ↓	12 ↓	2 ↑	1.8 ↑	0.8 ↓	Machado et al. (2014)
<i>Ascophyllum nodosum</i>	<i>In vivo</i>	Jersey cows	113 g d ⁻¹	11.5 ↓	–	–	–	–	Antaya et al. (2019)
<i>Saccharina latissima</i>	<i>In vivo</i>	Dairy cows	126–136 g d ⁻¹	ns	–	–	–	–	Muizelaar et al. (2023)
<i>Fucus serratus</i>	<i>In vivo</i>	Dairy cows	126–136 g d ⁻¹	ns	–	–	–	–	Muizelaar et al. (2023)

¹⁾ CH₄, methane; TVFA, total volatile fatty acids; C2, acetate; C3, propionate; C2/C3, the ratio of acetate to propionate; DMI, dry matter intake; Δ%: treatment response over control; ↑ and ↓ represent the percentage increase and decrease, respectively, of the value in the treatment group supplemented with seaweed compared with the control group. –, not quantified; ns, not statistically significant.

determined to be the key inhibitory component for enteric methanogenesis. Seaweeds with differing levels of bromoform will have distinct methane-inhibiting properties. A paper published by Thapa et al. (2020), found the genetic mechanism of biosynthesis of bromoform production in various seaweed. The key genes were *mbb1* and *mbb4* (vanadium-dependant haloperoxidases, VHPOs), with the highest bromoform production levels. The gene encoding reactive oxygen species-producing enzyme (*mbb2*) will be clustered with genes encoding bromoform-producing (Zhu et al. 2021). These results suggest that differences in bromoform-encoding genes contribute to disparities in the bromoform content of algae. However, knowledge of the bromoform biosynthesis routes, gene regulation, and catalytic activities remains to be resolved. More research is needed to determine the reasons for this discrepancy.

4. Challenges of using seaweeds in ruminant diet

4.1. Detrimental effect on intake, lactation performance, and rumen fermentation

In vivo studies showed decreased DMI by 10.8 to 14.2% in beef cattle (Kinley et al. 2020; Roque et al. 2021; Almeida et al. 2022) and 7 to 38% in dairy cows (Roque et al. 2019; Stefenoni et al. 2021; Angellotti et al. 2022; Wasson et al. 2022a) fed *A. taxiformis* and *A. armata*, respectively (Table 1). Energy-corrected milk (ECM) was decreased by 6 to 14% when *A. taxiformis* was fed to

lactating dairy cows (Stefenoni et al. 2021; Wasson et al. 2022a; Krizsan et al. 2023). However, Angellotti et al. (2022) found no difference in milk yield when *A. taxiformis* was supplemented at 0.3% OM of the diet, suggesting that dosage may not impact on milk production. In addition to milk production, Stefenoni et al. (2021) reported that contents in milk fat, milk true protein, and solid not fat percentages were not affected by the high inclusion level of *A. taxiformis*. However, when *A. taxiformis* was fed to dairy or beef cattle, increased concentrations of iodine and bromide in milk (Stefenoni et al. 2021) or meat (Roque et al. 2021) were observed. Conversely, the concentration of bromoform did not show a significant difference, which was consistent with the findings of Krizsan et al. (2023). Seaweeds are known to be accumulators of halogens, and previous studies have identified high levels of iodine in *A. taxiformis* (Mairh et al. 1989; Nunes et al. 2018). Consequently, high-dose feeding of this seaweed may pose a risk of elevated iodine residues in animal products. Although iodine consumption is necessary for humans, overconsumption is a health concern (Leung and Braverman 2014).

Methanogenesis prevents the accumulation of hydrogen in the rumen, which inhibits the activity of dehydrogenase (Martin et al. 2010). However, when methanogenesis was inhibited by seaweeds, methanogens were unable to consume the H₂, leading to a redirection of metabolic hydrogen toward propionate (H₂ sink) (Ungerfeld 2015). By summarizing previous studies in Table 1, we found a negative relationship between the methane reduction rate and the total VFA

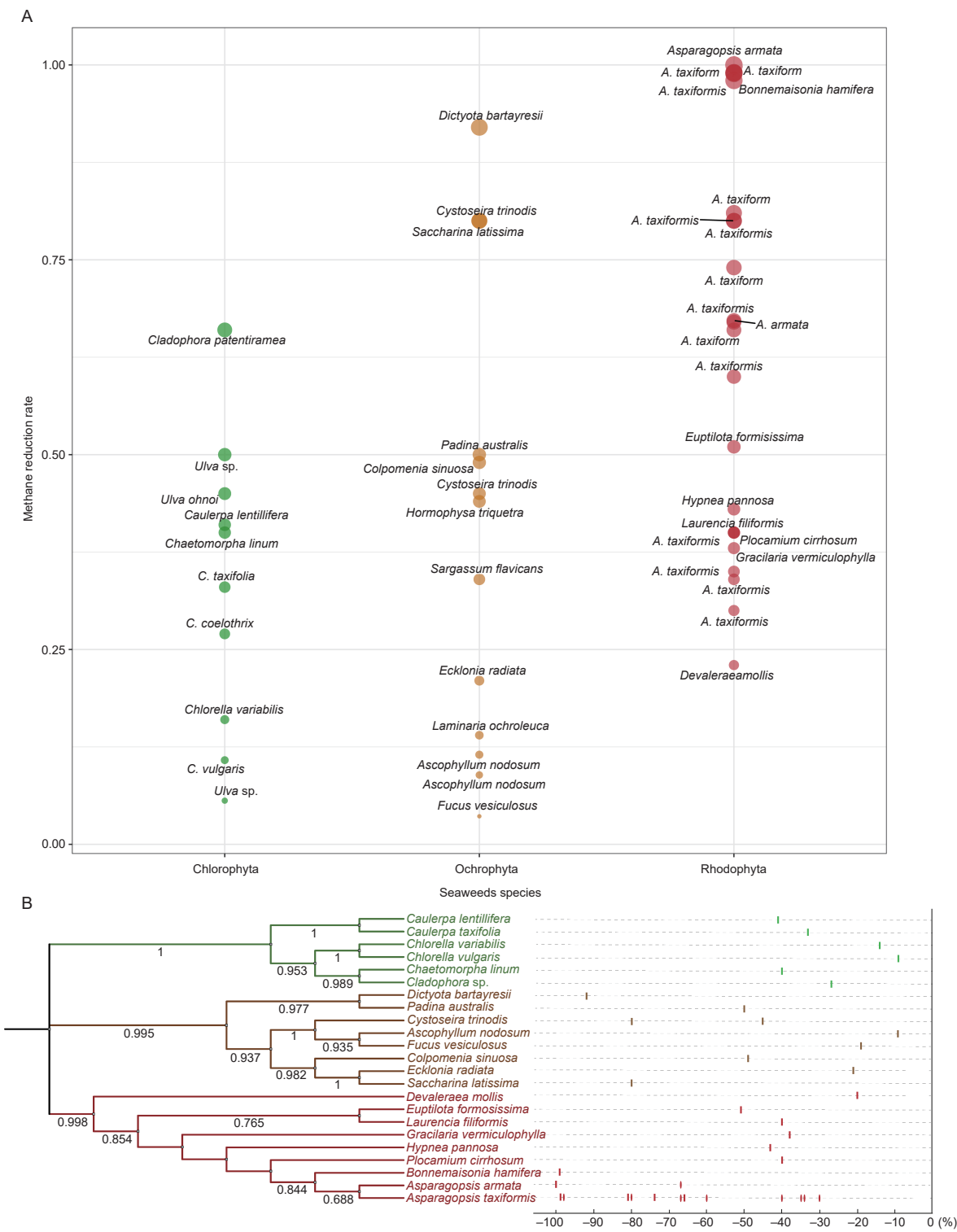


Fig. 3 The methane mitigation potential and evolutionary relationships of different seaweeds. A, summary of methane reduction rates of different seaweed species. According to seaweed species, this figure was divided into three sections (Chlorophyta, Ochrophyta and Rhodophyta). The size of the bubble represents the methane mitigation ability of different seaweeds. B, the phylogenetic tree was inferred by using the maximum likelihood method as can the General Time Reversible model. The tree with the highest log likelihood (−15,156.92) is shown. This analysis involved 23 nucleotide sequences. Codon positions included were 1st+2nd+3rd+Noncoding. There were a total of 2,181 positions in the final dataset. Evolutionary analysis was conducted in MEGA11. On the right side of the figure, the methane reduction abilities of various seaweed species are presented to evaluate the relationship between the phylogenetic proximity of the seaweeds and their potential for methane mitigation.

reduction rate (Fig. 4-A), and acetate reduction rate (Fig. 4-B). Conversely, the methane reduction rate was positively correlated with the propionate increase rate (Fig. 4-C), suggesting that reducing methane emissions by seaweeds may redirect metabolic hydrogen in the rumen to propionate production.

4.2. Potential bromoform residue in products

According to research by the Environmental Protection Agency (EPA) (EP 2015), nervous system injury due to high levels of bromoform inhalation or ingestion in humans and animals leads to brain malfunction and liver and

Table 3 Summary of studies using green algae (Chlorophyta) for methane reduction, including methods, inclusion level, results, and references¹⁾

Seaweed	<i>In vivo</i> / <i>In vitro</i>	Dose	CH ₄ (Δ%)	TVFA (Δ%)	C2 (Δ%)	C3 (Δ%)	C2/C3 (Δ%)	References
<i>Caulerpa lentillifera</i>	<i>In vitro</i>	160 mg OM	41 ↓	–	–	–	–	Dubois et al. (2013)
<i>Caulerpa taxifolia</i>	<i>In vitro</i>	0.2 g OM	33 ↓	17 ↑	4.8 ↑	8.9 ↓	17 ↑	Machado et al. (2014)
<i>Cladophora coelothrix</i>	<i>In vitro</i>	0.2 g OM	27 ↓	1 ↓	0.3 ↓	4.9 ↑	6.3 ↓	Machado et al. (2014)
<i>Cladophora patentiramea</i>	<i>In vitro</i>	0.2 g OM	66 ↓	9 ↓	0.2 ↓	4.9 ↑	6.3 ↓	Machado et al. (2014)
<i>Chaetomorpha linum</i>	<i>In vitro</i>	0.2 g OM	40 ↓	ns	2.7 ↓	13 ↑	14 ↓	Machado et al. (2014)
<i>Chlorella vulgaris</i>	<i>In vitro</i>	25% DM	10.8 ↓	7 ↓	21 ↓	30 ↑	–	Sucu (2020)
<i>Chlorella variabilis</i>	<i>In vitro</i>	25% DM	16 ↓	13 ↓	25 ↓	17 ↑	–	Sucu (2020)
<i>Uva</i> sp.	<i>In vitro</i>	0.2 g OM	50 ↓	ns	0.8 ↓	4.5 ↑	5.5 ↓	Machado et al. (2014)
<i>Ulva</i> sp.	<i>In vitro</i>	10% OM	5.6 ↓	6 ↓	0.1 ↑	0.5 ↓	–	Mihaila et al. (2022)
<i>Ulva ohnoi</i>	<i>In vitro</i>	0.2 g OM	45 ↓	6 ↓	2.9 ↑	4.2 ↓	6.3 ↑	Machado et al. (2014)

¹⁾ CH₄, methane; TVFA, total volatile fatty acids; C2, acetate; C3, propionate; C2/C3, the ratio of acetate to propionate; DMI, dry matter intake; Δ%, treatment response over control; ↑ and ↓ represent the percentage increase and decrease, respectively, of the value in the treatment group supplemented with seaweed compared with the control group. –, not quantified; ns, not statistically significant.

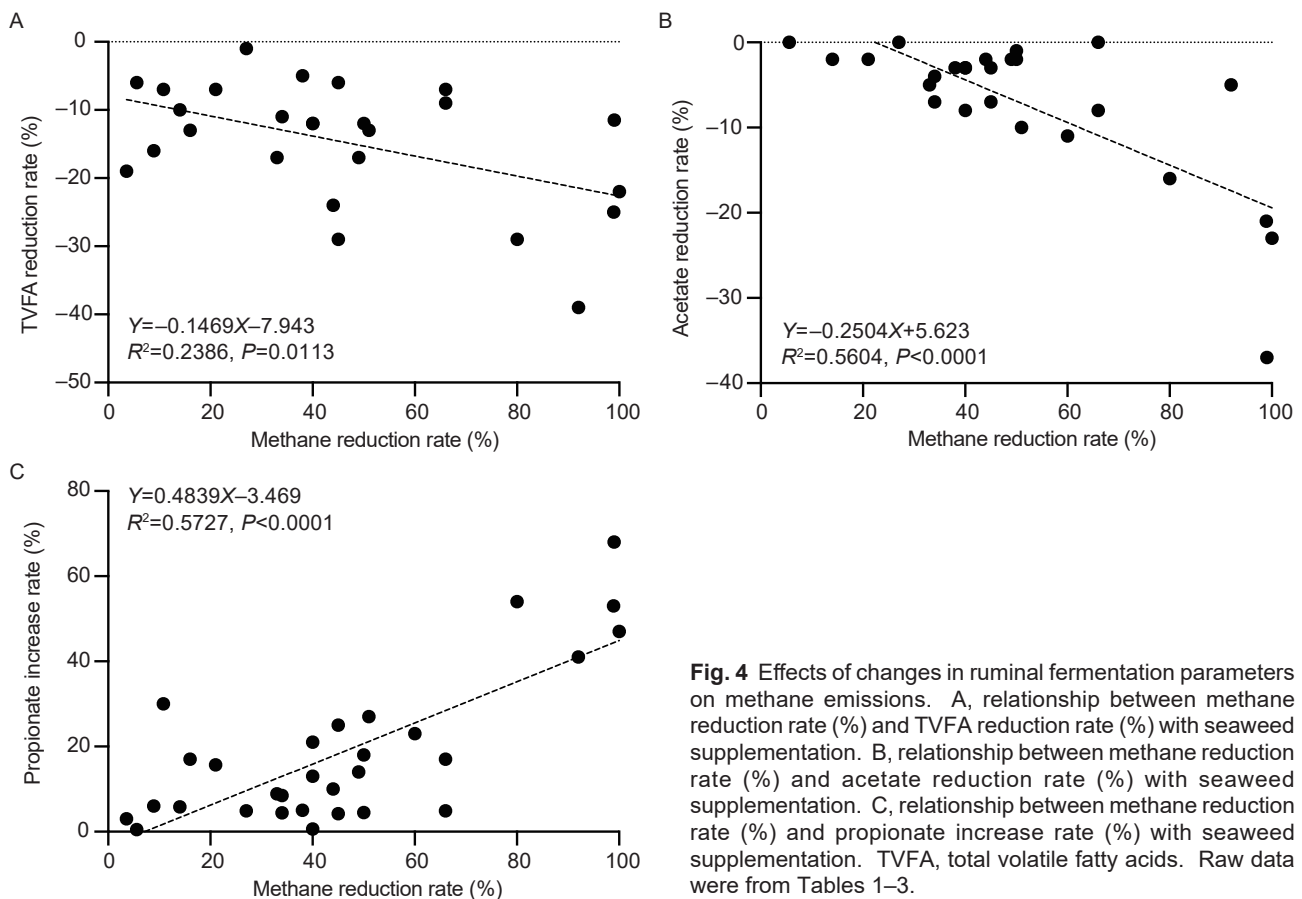


Fig. 4 Effects of changes in ruminal fermentation parameters on methane emissions. A, relationship between methane reduction rate (%) and TVFA reduction rate (%) with seaweed supplementation. B, relationship between methane reduction rate (%) and acetate reduction rate (%) with seaweed supplementation. C, relationship between methane reduction rate (%) and propionate increase rate (%) with seaweed supplementation. TVFA, total volatile fatty acids. Raw data were from Tables 1–3.

kidney damage. Therefore, the EPA has set a limit of $80 \mu\text{g L}^{-1}$ for bromoform consumption in drinking water. The bromoform concentrations in milk produced by dairy cows fed low (0.5%) and high (1%) *A. taxiformis* levels were 0.11 to $0.15 \mu\text{g L}^{-1}$, which is over 500 times lower than the maximum standard, according to Roque *et al.* (2019). Similarly, Kinley *et al.* (2020) found no changes in the meat quality of beef cattle supplemented with *A. taxiformis* at a concentration of 0.00, 0.05, 0.10, and 0.20% of feed OM. Romero *et al.* (2022) showed that bromoform in *A. taxiformis* was degraded quickly during incubation, with 70% degradation in the first 30 min and nearly 90% after 3 h. After 12 h of incubation, the bromoform was not detectable. However, Muizelaar *et al.* (2021) found that although bromoform did not accumulate in animal tissue, it was excreted in urine and milk and caused abnormalities of rumen wall papillae in lactating dairy cows fed *A. taxiformis*. According to Li *et al.* (2018), no detectable levels ($<0.05 \text{ mg kg}^{-1}$) of bromoform or dibromochloromethane were identified in any muscle or fat sample of sheep. However, five of ten sheep offered *A. taxiformis* had the mucosal lining of the rumen floor examined grossly after death, showing a variable-sized area of nodular proliferation, white-tan discoloration, and blunted ruminal papillae. The effects of seaweeds on the safety of ruminant products need further evaluation. Most countries or regions are currently cautious about using seaweeds as feed additives to reduce enteric methane emissions from ruminants. So far, only the California Department of Food and Agriculture has approved one of the most effective seaweeds for cattle as an inhibitor. Moreover, a recent report from Sweden's EPA proposes immediate government action to encourage the adoption of these feed additives through incentives.

4.3. Storage of bromoform in seaweeds

The efficacy of seaweeds in mitigating enteric methane heavily depends on the processing and storage of bioactive compounds within their gland cell walls (Paul *et al.* 2006). Vucko *et al.* (2017) utilized a factorial design to determine the effects of rinsing (unrinsed, dip rinsed, and submerged), freezing (frozen and not frozen), and drying (freeze-dried, kiln-dried, and dehydrated) of *A. taxiformis* on methane production *in vitro* and concentration of bromoform in the biomass. The results indicated that freeze-drying was the most effective processing method for maintaining bromoform concentration and anti-methanogenic activity. Stefenoni *et al.* (2021) evaluated the effect of storage time, temperature, and light exposure on the concentration of bromoform in *A. taxiformis*, and

found that after 4 mon of storage, the concentration of bromoform decreased by 75% when stored in the dark and by 84% when stored under luminescent light, while storage temperature did not affect the concentration of bromoform. Magnusson *et al.* (2020) demonstrated that immersion of *Asparagopsis* genus biomass in oil could maintain high concentrations of bromoform for at least 12 wk, even when stored at room temperature (25°C), while water immersion resulted in 40% less bromoform extraction and at least 47% loss after 12 wk of storage. Tan *et al.* (2022) investigated the relatively long-term storage of Freeze-Dried-Asp and Asp-Oil under different storage conditions (different temperatures, presence of fluorescent light, and exposure to open air). Asp-Oil was a more desirable method of preserving bromoform compared to Freeze-Dried-Asp. Freeze-Dried-Asp was only stable at -20°C , whereas Asp-Oil had a higher or no change in bromoform concentrations after 24-wk storage at 40, 25, 4 and -20°C in dark conditions (Tan *et al.* 2022). We have concluded that the storage and stability of bromoform in seaweed are influenced by factors such as cultivation, processing, and transportation. This poses challenges and shortcomings in establishing a commercial supply chain for seaweed utilization. In addition, more studies are needed to explore the influence of factors such as harvest time and regional differences on bromoform concentration in seaweeds.

4.4. Potential environmental impact of bromoform in seaweeds

In addition to potential damage to animal health, bromoform can catalytically destroy ozone in the troposphere and stratosphere via chemical reactions (WMO 2018). While large-scale *A. taxiformis* farming is one of the most promising options to produce anti-methanogenic feed additives, how *A. taxiformis* farming may affect the atmospheric environment needs to be investigated and elucidated. A recent study by Jia *et al.* (2022) evaluated various cultivation scenarios to determine the potential risks of bromoform released from *A. taxiformis* farming near Australia on the stratospheric ozone layer. The study showed that under typical operating conditions, the planned operation of *Asparagopsis* seaweed cultivation farms at Triabunna, Yamba, Geraldton, and Darwin, with an annual yield of 34,674 mg dry weight, does not have an impact on the ozone layer. In addition, according to Glasson *et al.* (2022), bromoform from cultivated *A. taxiformis* biomass had little impact on the overall natural and anthropogenic atmospheric loading. More studies are needed to verify if large-scale artificial cultivation of

seaweed could potentially cause detrimental effects on ozone.

4.5. Costs of seaweed farming and processing

Asparagopsis taxiformis farming was still in its infancy, with only some startup companies including VOLTA, Greener Grazing, Sea Forest, Symbrosia, and Blu Ocean Barns. Commercial-scale production of seaweeds is still not available, and dried *A. taxiformis* is valued between US\$100 to 146 kg⁻¹ (Jayvee et al. 2021). The cost of wild *A. taxiformis* harvested in Australia was even higher, around US\$200 kg⁻¹ (Cotter et al. 2015). Even if the price was reduced to less than US\$5 kg⁻¹, adding 20 kg of dried *A. taxiformis* would cost an additional US\$100 per tonne of diet. Therefore, the cost of meat and milk products from cows consuming such a diet would increase roughly 20% (Trevor Whittington 2022). As the main functional component, bromoform could be easily volatilized during preservation and processing. To avoid volatilization, vegetable oil was used (Magnusson et al. 2020) which would potentially increase the cost. Another important step, freeze-drying, would also increase the processing cost of seaweeds.

5. Recommendations for future research

5.1. Exploration of more seaweed resources

Although there are more than 12,000 species of seaweeds in the world, only 221 are considered commercially valuable, and even fewer species are cultivated intensively (FAO and WHO 2021). *Asparagopsis taxiformis* species appear to be the best candidate as a feed additive due to their effectiveness at low inclusion rates (Wasson et al. 2022b). Other species with low bromoform concentrations, such as *Bonnemaisionia hamifera* (Mihaila et al. 2022), *Dictyota bartayresii* (Machado et al. 2014), and *Cystoseira trinodis* (Dubois et al. 2013) could also reduce methane emissions at a comparable level of inclusion. This research was mainly conducted in coastal countries such as Australia, Canada, Norway, and Denmark. The United States, the United Kingdom, the Netherlands, and New Zealand have also begun exploring the potential of various seaweeds to reduce methane emissions in recent years (Fig. 5). However, almost all of the seaweeds consumed worldwide are produced in Asia, with 57 percent being produced in China alone, followed by Indonesia, the Republic of Korea, and the Philippines (Nin-Pratt et al.



审图号：[GS(2016) 2958]

Fig. 5 The country's distribution of seaweed is used as a feed additive to mitigate methane emissions before 2024. Countries highlighted in yellow on the global map have published research on seaweed as a methane inhibitor.

2022). There are still many unexploited seaweed resources in Asian countries. Additional research is still required to evaluate the potential of novel types of seaweeds for methane mitigation in ruminants.

5.2. Identification of bioactive compounds other than bromoform in seaweeds

Halogen compounds, including bromoform, inhibit the ability of methanogen enzymes to produce methane by reacting with vitamin B₁₂ (Bauchop 1967). However, other bioactive compounds in seaweeds, including protein hydrolysates, polysaccharides, phlorotannins, and alkaloids (Holdt and Kraan 2011; Hafting *et al.* 2015) may also play a role in reducing methane emissions. Phlorotannins derivatives (phlorofucofuroeckol-A, dieckol, and bieckol), a class of polyphenolic compounds found in brown algae, may reduce fiber digestibility and alter rumen fermentation, resulting in decreased H₂ production (Wang *et al.* 2008). According to Martinez-Fernandez *et al.* (2017), phloroglucinol degradation in the rumen promotes the capture of excess H₂ generated from methanogenesis inhibition. Furthermore, Kim *et al.* (2022) discovered that supplementation with phlorotannins reduced methane emissions without impairing rumen fermentation characteristics. In addition, saponin is known to be a natural detergent with membrane-degrading groups complex with sterols in protozoal cell membranes, which may reduce the population of protozoa in rumen fluid and change ruminal fermentation (Patra and Saxena 2009). Protozoa-associated methanogens are among the most active communities involved in rumen methanogenesis. Reducing protozoa could reduce methanogens, thereby reducing methane emissions (Patra *et al.* 2017). Alkaloids and fucoidans have antibacterial potential (Pérez *et al.* 2016); however, their efficacy on methane mitigation needs further exploration.

In addition to the bioactive components, rumen metabolites potentially inhibit methane emissions when seaweeds are fed to ruminants and metabolized by rumen microorganisms. Olijhoek *et al.* (2022) showed that supplementation of Nordic macroalgae *Saccharina latissimi* could reduce methane production by up to 80% *in vitro* without affecting the degradability of the substrate. Notably, halomethanes were not detected in the *S. latissimi* samples by Gas Chromatography-Mass Spectrometry (Olijhoek *et al.* 2022). In addition, the second principal component (PC2) revealed that several potential compounds, including 'Compound X', were able to reduce methane emissions by 47 and 79%, respectively, with medium dose and high dosage *in vitro* (Thorsteinsson *et al.* 2022). Future research is needed

to explore more seaweed compounds that potentially mitigate methane emissions.

5.3. Combination of anti-methanogenic inhibitors

Among the most successful anti-methanogenic compounds tested were coenzyme M analogues (BES, CES, and BPS), and HMAs (chloroform, bromoform, dibromochloromethane, and bromochloromethane), which reduced methane production by up to 50% (Bauchop 1967; Tomkins *et al.* 2009; Liu *et al.* 2011; Abecia *et al.* 2012). In addition, 3-NOP acts on MCR, a key enzyme of the methanogenesis pathway, which is a terminal process used by methanogenesis to produce methane (Kebreab *et al.* 2022). These small molecular compounds can mitigate methane production by inhibiting key enzymes in methanogenesis, when combined the mitigation effect may be stronger. In the rumen, methanogenesis is considered to be the primary pathway for H₂ removal. Several alternative H₂ sinks have been identified when rumen methanogenesis is inhibited, including fumarate, nitrate, sulfate, and homoacetogenesis (Greening *et al.* 2019; Karekar *et al.* 2022). Microbial reduction of nitrates and sulfate to ammonia and sulfide provides a thermodynamically more favorable electron and H₂ sink than methanogenesis (Ungerfeld 2020). Fumarate reduction is also thermodynamically more favorable than the reduction of carbon dioxide to methane (Kohn and Boston 2000). In addition, Homoacetogens have the ability to use gaseous substrates to produce acetic acid (Karekar *et al.* 2022) under normal conditions, the homoacetogenic process can be carried out spontaneously. However, this process only occurs when H₂ partial pressure is high and the homoacetogenic bacteria are in suitable conditions to maintain the ecological environment. In general, homoacetogens are not prevalent when sufficient methanogens are present in the ecosystem. If the methanogenesis process is inhibited, homoacetogens can use excess H₂ to produce acetic acid (Karekar and Ahring 2023). The use of combinations of seaweeds, as well as additional H₂ sinks, which can mitigate methane emission without compromising rumen fermentation functions, should be developed.

6. Conclusion

Seaweeds were proven to have significant methane mitigation potential. However, there are still many limitations, including the potential detrimental effect of high bromoform content on animal health as well as the residuals in products, lack of commercially available resources that may enhance the cost, unexplored novel

species, and bioactive compounds that may also inhibit methanogenesis other than bromoform. Future studies are needed to explore the role of seaweeds play in modulating rumen methanogenesis with the help of integrated omics, which are essential to develop better strategies to maximize the efficacy of seaweeds in methane mitigation.

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Declaration of competing interest

The authors declare that they have no conflict of interest.

Ethical statements

This research did not involve animal ethics experiments.

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